

## Biodegradation of Para-Amino Acetanilide by *Halomonas* sp. TBZ3

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**Background:** Aromatic compounds are known as a group of highly persistent environmental pollutants. *Halomonas* sp. TBZ3 was isolated from the highly salty Urmia Lake of Iran. In this study, characterization of a new *Halomonas* isolate called *Halomonas* sp. TBZ3 and its employment for biodegradation of para-amino acetanilide (PAA), as an aromatic environmental pollutant, is described.

**Objectives:** This study aimed to characterize the TBZ3 isolate and to elucidate its ability as a biodegradative agent that decomposes PAA.

**Materials and Methods:** Primarily, DNA-DNA hybridization between TBZ3, *Halomonas denitrificans* DSM18045T and *Halomonas saccharovitans* LMG 23976T was carried out. Para-amino acetanilide biodegradation was assessed using spectrophotometry and confirmed by gas chromatography-mass spectroscopy (GC-MS). Parameters effective on biodegradation of PAA were optimized by the Response Surface Methodology (RSM).

**Results:** The DNA-DNA hybridization experiments between isolate TBZ3, *H. denitrificans* and *H. saccharovitans* revealed relatedness levels of 57% and 65%, respectively. According to GC-MS results, TBZ3 degrades PAA to benzene, hexyl butanoate, 3-methyl-1-heptanol and hexyl hexanoate. Temperature 32.92°C, pH 6.76, and salinity 14‰ are the optimum conditions for biodegradation with a confidence level of 95% (at level  $\alpha = 0.05$ ).

**Conclusions:** According to our results, *Halomonas* sp. TBZ3 could be considered as a biological agent for bioremediation of PAA and possibly other similar aromatic compounds.

**Keywords:** Biodegradation; *Halomonas*; Waste Management; Bioremediation; Environmental Pollution

### 1. Background

Aromatic compounds are known as a group of highly persistent environmental pollutants (1), most of which possess toxic, carcinogenic and mutagenic properties (2). These chemicals are also recalcitrant to degradation. For example, the half-life of the hydrolyzed form of the dye compound, reactive blue 19, is about 46 years at pH 7 and 25°C (3). Para-amino acetanilide (PAA) is an aromatic compound used in the color industry; it is used as an intermediate in the manufacture of azo dyes and pharmaceuticals (4). Several physicochemical and biological methods are available for the removal of aromatic pollutants from the environment. Physicochemical approaches such as filtration, coagulation, activated carbon, photo-degradation, and chemical flocculation are used for eliminating aromatic compounds from the environment. However, physicochemical methods of removing aromatic compounds from the environment have some disadvantages such as

high costs, being coupled with sludge formation, and release of toxic materials (5). Furthermore, the traditional techniques applied in textile wastewater treatment, such as chemical coagulation/flocculation, membrane separation or activated carbon adsorption are known to be costly and only transfer the pollutants from the liquid to the solid phase (6). Therefore, bioremediation is a cost-effective and environmental-friendly alternative for physical and chemical methods of cleaning environmental pollutants. Bioremediation (which is largely dependent on biotic and abiotic factors such as bacterial population size, pH, temperature, etc.) (7, 8) has been intensively studied over the past two decades, driven by the need for a low-cost, sustainable with the natural environment (9).

Among all organisms, microorganisms, especially bacteria, have been widely studied and frequently used for biodegradation strategies; since they are able to use

chemical pollutants as the sole source of carbon and nitrogen (1). Some research groups have reported the biodegradation of aromatic compounds by bacteria belonging to *Halomonas* species (10, 11). The genus *Halomonas* presents a group of Gram-negative, rod-shaped and aerobic halotolerant bacteria, which are able to grow in a wide range of salt concentrations (12). Bacteria belonging to genus *Halomonas* have applicable potentials in various fields of industry, ecology and biotechnology (13, 14). Over the past decade, interest in *Halomonas* species has also focused on their ability to degrade aromatic compounds such as benzoic, salicylic, p-hydroxybenzoic, p-coumaric, cinnamic, phenylacetic and ferulic acid (15). Oie et al. (11) showed that the haloalkaliphilic bacterium, *Halomonas campisalis*, is able to degrade model aromatic compounds; benzoate and salicylate, at pH 9 and in up to 100 g/L NaCl. They detected metabolites, such as catechol and cis-muconate.

Located northwest of Iran, in the Azarbayjan region, Urmia Lake is the largest saline lake in the Middle East and the second largest salt water lake on the earth. Urmia Lake, like the dead-sea, is well known for its extreme water salinity. Traditionally, the lake water is believed to possess healing properties such as curing rheumatism. The NaCl concentration of Urmia Lake was determined to be about 34 g/L in 1915, which has increased to more than 300 g/L because of drought, evaporation, and higher demands for agricultural water in the lake's basin. This continuing development has simultaneously become a major concern and the survival of the lake has been a critical and challenging issue in the recent years. Such an evolution encouraged us to study the halophilic bacterial population of Urmia Lake resulting in isolation of some new *Halomonas* isolates including *Halomonas* sp. TBZ3 (16).

## 2. Objectives

Considering the environmental problems posed by aromatic compounds such as PAA (used by various factories in the area) and the ability of *Halomonas* bacteria to degrade aromatic compounds (10, 11), this study aimed to characterize the TBZ3 isolate, and to elucidate its ability as a biodegradative agent to decompose PAA. Additionally, some factors affecting biodegradation rate (temperature, pH and salinity) were optimized using the response surface methodology (RSM) method.

## 3. Materials and Methods

### 3.1. Description of *Halomonas* sp. TBZ3

*Halomonas* sp. TBZ3 was isolated from water samples of the Urmia Lake. Based on 16S rRNA gene sequence (EU305729 accession number), TBZ3 belongs to the *Halomonas* genus. These bacteria are Gram-negative, oxidase- and catalase-positive. Its colonies are convex,

smooth and translucent creamy. Growth occurs in a medium with 7.5% - 20% (w/v) NaCl. No growth occurs in the absence of salt. It grows at temperatures between 10 and 45°C and in pH ranging from 6 to 9. It does not produce acid from D-glucose, D-mannose, D-fructose, maltose, D-mannitol, rhamnose, sucrose, D-trehalose, D-galactose, cellobiose, D-raffinose, xylose, lactose, ribose, arabinose, D-salicin and aesculin. Indole and urease activities, and H<sub>2</sub>S production are negative. Phenylalanine-deaminase production is positive. Tween 20 and Tween 80 are hydrolyzed, yet tyrosine, starch and casein are not hydrolyzed. Nitrate is not reduced to nitrite. Arginine is not deaminated. In order to further characterize and confirm that the TBZ3 isolate is not identical with other known bacterial type strains; the isolate was subjected to complementary characteristic studies (16).

### 3.2. DNA-DNA Hybridization and DNA G + C Content

Primarily, DNA-DNA Hybridization (DDH) between TBZ3 and reference isolate *Halomonas denitrificans* DSM18045T, was carried out as described by De Ley et al. (17) with the modifications described by Huss et al. (18), by German collection of microorganisms and cell cultures (DSMZ). The DNA-DNA hybridization between TBZ3 and *Halomonas saccharovitans* LMG 23976T was done by a Belgian coordinated collection of microorganisms, laboratory of Microbiology of Gent (BCCM/LMG), Gent, Belgium. Hybridization was performed in the presence of 50% formamide at 50°C, according to a modified version of the method described by Ezaki et al. (19).

### 3.3. Culture Media and Aromatic Compounds

Para-Amino Acetanilide was provided by the Boyakh Saz Company. At first, the previously isolated TBZ3 cells were cultured in marine Broth 2216 (Difco) and incubated at 30°C to reach a suitable state of growth. Then, 400 µL aliquots of fresh culture (OD 0.5) were inoculated in Mineral Medium (MM) pH 7.0 to conduct the biodegradation process. This medium contained (per liter): K<sub>2</sub>HPO<sub>4</sub> (8.0 g), KCl (1.0 g), NH<sub>4</sub>Cl (1.5 g), MgSO<sub>4</sub> (0.2 g) and NaCl (70.0 g), and was supplemented with 40 ppm of filter sterilized (0.45 µm) PAA as the sole source of carbon (20, 21). Biodegradation tests were conducted in series consisting of, I) MM + PAA + bacterial isolate; II) MM + PAA, and III) medium + bacterial isolate; with (II) and (III) samples serving as controls (22).

### 3.4. Para-Amino Acetanilide Biodegradation Capacity of *Halomonas* sp. TBZ3

To evaluate the PAA degradation ability of *Halomonas* sp. TBZ3, this substrate was provided with a concentration of 40 ppm as the sole source of carbon. Then, the samples were incubated in a 100 rpm Vision (VS-8480) shaking incubator at 30°C. Screening was done through

spectroscopic scanning of the samples using a spectrophotometer (UNICO UV-1200) at wavelengths of 200 - 800 nm with  $\lambda_{\max}$  of 258 nm. Before starting this spectroscopic operation, samples were centrifuged at 13000 rpm for 15 minutes. The spectrophotograms of the samples before and after the incubation were compared. All experiments were done at least in triplicates (23).

### 3.5. Quantitative Estimation of the Process

Quantitative estimation was accomplished for PAA bacterial biodegradation. Para-Amino Acetanilide was fed to the bacterial isolate, at a concentration of 40 ppm, as the sole source of carbon. The absorbance was measured at 258 nm using the UNICO VU-1200 spectrophotometer. The absorbance was used as a measure of biodegradation rate. The biodegradation rate was estimated using Equation 1.

$$(1) \quad D(\%) = (A_0 - A) \times \frac{100}{A_0}$$

Where,  $A_0$  is the absorbance value of the control sample,  $A$  is the absorbance value after degradation, and  $D$  (%) is the biodegradation percentage (23).

### 3.6. Biodegradation Metabolites Detection

Detection of metabolites produced during PAA biodegradation was carried out using gas chromatography-mass spectroscopy (GC-MS). Analysis of GC-MS was conducted by GC-MS Agilent 6890 (USA) equipped with a 30 m  $\times$  0.25 mm  $\times$  25  $\mu$ m HP-5MS capillary column coupled with an Agilent 5973 mass spectrometer (Agilent Technologies, Palo Alto, CA) operating in EI mode at 70 eV with the following features: helium as the carrier gas with a pressure of 34 psi at injection port and a quadrupole filter. The injection port temperature was 280°C. The initial oven temperature was 50°C and post-run temperature was 300°C (21, 24).

### 3.7. Optimization of Biodegradation Conditions

The RSM method was exploited to optimize the biodegradation conditions and for this purpose, Minitab 15 software was employed. Three independent variables (pH, temperature and concentration of salt) were selected as they were expected to affect biodegradation efficiency. Eight ( $2^3$ ) factorial points and six ( $2 \times 3$ ) axial points with six replicates at the center point, thus a total of 20 experiments, were conducted in a randomized order. All of the experiments were done at least in triplicate. The center point replicates were selected to verify any change in the assessment procedure, as a measure of exactitude property. A full quadratic model for this design is given as Equation 2:

$$(2) \quad D = \beta_0 + \sum_j^k \beta_j x_j + \sum_j^k \beta_{jj} x_j^2 + \sum_i \sum_{<j} \beta_{ij} x_i x_j + \epsilon$$

Where,  $D$  is the degradation efficiency,  $x_i$  and  $x_j$  are variables ( $i$  and  $j=1$  to  $k$ ),  $\beta_0$ ,  $\beta_j$ ,  $\beta_{jj}$  and  $\beta_{ij}$  are the constant term, coefficients of linear, quadratic and second-order terms, respectively,  $\epsilon$  is the error, and  $k$  is the number of independent variables (here = 3). Each variable was coded at five levels between -2 and + 2, according to the Equation 3 (25).

$$(3) \quad x_i = \frac{X_i - X_0}{\delta X}$$

Where,  $x_i$  is the coded value of the variable,  $X_i$  is the existent value of the variable,  $X_0$  is the center point value, and  $\delta X$  is the step change between the levels. Following initial investigations and determination of efficient parameters in the biodegradation procedure, the model was expressed as Equation 4 (25).

$$(4) \quad D = 0.33837 + 0.01072T + 0.00651pH + 0.02948C - 0.05450T^2 - 0.03562pH^2 + 0.00523C^2 - 0.03878T.pH + 0.01752T.C - 0.01032pH.C$$

At confidence level of 95%, only the linear terms,  $T^2$  and  $T$ . pH had a meaningful effect on the results. Therefore, considering this point, the optimum conditions (for the maximum biodegradation efficiency) were predicted by the software. Five levels of temperature, pH and salinity were selected to reach the optimum values for the biodegradation process. These values for each variable were selected based on primary results gained from preliminary experiments and the reports of related scientific literature. Table 1 demonstrates the experimental design procedure.

**Table 1.** Conditions for Each Biodegradation Experiment Run, Designed by the Response Surface Methodology Method <sup>a</sup>

Standard Order	Temperature, °C	pHZ	C, % w/v
1	25 (-1)	6.5 (-1)	3.5 (-1)
2	35 (1)	6.5 (-1)	3.5 (-1)
3	25 (-1)	7.5 (1)	3.5 (-1)
4	35 (1)	7.5 (1)	3.5 (-1)
5	25 (-1)	6.5 (-1)	10.5 (1)
6	35 (1)	6.5 (-1)	10.5 (1)
7	25 (-1)	7.5 (1)	10.5 (1)
8	35 (1)	7.5 (1)	10.5 (1)
9	20 (-2)	7.0 (0)	7.0 (0)
10	40 (2)	7.0 (0)	7.0 (0)
11	30 (0)	6.0 (-2)	7.0 (0)
12	30 (0)	8.0 (2)	7.0 (0)
13	30 (0)	7.0 (0)	0.0 (-2)
14	30 (0)	7.0 (0)	14.0 (2)
15 - 20	30 (0)	7.0 (0)	7.0 (0)

<sup>a</sup> The values between parenthesis are related to coded values.

Maximum absorbance at 258 nm was evaluated under these conditions using a spectrophotometer for measuring biodegradation rate. Response Surface Methodology is used to optimize more than one parameter simultaneously; parameters related to the biodegradation process are optimized one by one. For instance, Dehghani et al. (26) optimized pH in alachlor biodegradation. Response Surface Methodology naturally includes a proper number of repetitions, yet to ensure accuracy, all of the optimization experiments were done in triplicates (10, 27).

## 4. Results

### 4.1. *Halomonas* sp. TBZ3 DNA G + C Content and DNA-DNA Relatedness

The G + C content of the genomic DNA of *Halomonas* sp. TBZ3 was 67.1 mol%. The DNA-DNA hybridization experiments between *Halomonas* sp. TBZ3, *H. denitrificans* DSM 18045T and *H. saccharovitans* LMG 23976T revealed relatedness levels of 57% and 65%, respectively.

### 4.2. Quantitative Examination of Biodegradation

Aromatic compounds are stable chemicals and are slowly degraded in the environment. However, biodegradation of these compounds using *Halomonas* species has already been shown (10, 15). Accordingly, primary investigation of PAA biodegradation by the TBZ3 isolate was conducted. To achieve this, PAA was added to the culture medium as the sole carbon source and its biodegradation was measured using spectrophotometric techniques. Spectrophotometric analysis showed that isolate TBZ3 is able to degrade PAA. Comparative results indicated that TBZ3-mediated biodegradation rate of PAA was 32.91% after a one-week incubation period.

### 4.3. Biodegradation Metabolites Detection

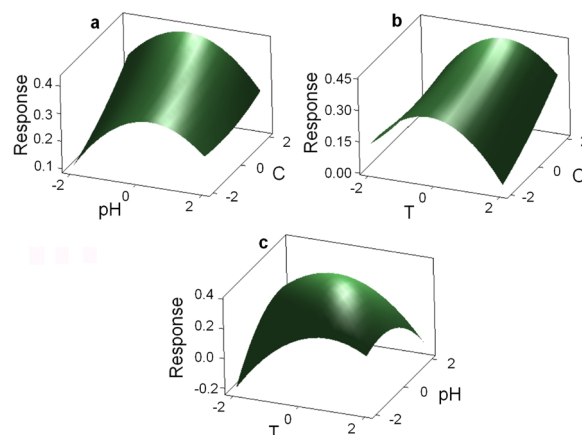
Detection of PAA biodegradation products was done by the GC-MS method. The GC-MS results showed that PAA is degraded by TBZ3. Furthermore, GC-MS revealed products at retention times of 2.17, 4.65, 4.88, 5.41 and 7.01 minutes. Mass spectroscopy data demonstrated that the biodegradation metabolites of PAA by TBZ3 were benzene, hexyl butanoate, 3-methyl-1-heptanol and hexyl hexanoate. The results of GC-MS analysis are summarized in Figure 1.

### 4.4. Optimization of Biodegradation Conditions

Response Surface Methodology was applied to find the optimum conditions for biodegradation temperature, pH, and salinity percentage. The spectrophotometric curves of TBZ3-treated PAA samples and controls were compared. The response surface plots are shown in Figure 2. The results showed that PAA is biodegraded by isolate TBZ3. It was found that temperature of 32.92°C, pH 6.76 and salinity of 14% were the optimum conditions for PAA biodegradation.

No.	Compound Name	Structure	Retention Time, min	Main Fragments
1	PAA		5.41	82, 109, 150, 54
2	Benzene		2.17	77
3	hexyl butanoate		4.65	89, 129, 172, 43
4	3-methyl-1-heptanol		4.88	112, 84, 70, 55
5	hexyl hexanoate		7.01	43, 84, 117

**Figure 1.** Gas Chromatography-Mass Spectroscopy Identified Products for Biodegradation of Para-Amino Acetanilide



**Figure 2.** Response surface plot of a: pH versus salinity percentage (C) at temperature of 30°C, b: temperature (T) versus salinity percentage (C) at pH = 7, c: temperature (T) versus pH at 7% salinity

## 5. Discussion

Bioremediation provides a low-cost and environmentally safe way to remove pollutants from the environment (8). Aromatic compounds are stable chemicals and are slowly degraded in the environment. Biodegradation of these compounds using *Halomonas* species has already been shown (11, 15). In this study, a native, newly isolated *Halomonas* strain was exploited for biodegradation of an aromatic compound used in the area. Similar studies have used native bacteria for bioremediation goals. For example, Moghadam et al. (8) used bacteria isolated from Nayband Bay region for biodegradation of phenanthrene. In another study, Abari et al. (7) introduced a novel bacterial isolate capable of toluene biodegradation. Taking into account that relatedness levels significantly below the threshold value of 70% are generally accepted for species delineation (28, 29), the DDH values of TBZ3 justify distinction of strain TBZ3 from *H. denitrificans* DSM18045T and *H. saccharovitans* LMG 23976T at the species level.

Mass spectroscopy data helped us identify the biodegradation metabolites of PAA by TBZ3. In some studies, GC-MS has been applied to determine metabolites produced during biodegradation of aromatic compounds. For example, Lin et al. (21) identified degradation metabolites

of naphthalene by isolate *Bacillus fusiformis* using GC-MS. Khataee et al. (29) also used the GC-MS method for detection of metabolites produced during the degradation of basic red 46. Abari et al. (7) used gas chromatography as a sensitive and accurate method to follow up toluene biodegradation rate by bacteria isolated from wastewater, and another study by Moghadam et al. used the same method to show that phenanthrene can be degraded by bacteria isolated from coastal sediments of Nayband bay in Iran (8). Apart from revealing aromatic compound degradation capability, biodegradation metabolites were also detected in this study (such as benzene, hexyl butanoate, 3-methyl-1-heptanol and hexyl hexanoate).

According to RSM results, temperature of 32.92°C, pH 6.76, and salinity of 14% were the optimum conditions for biodegradation of PAA by *Halomonas* sp. TBZ3 with a confidence level of 95% (at level  $\alpha = 0.05$ ). The experiments showed that biodegradation of PAA is efficiently increased under these optimum conditions. After optimization, biodegradation rate rises from 32.91% to about 66%. In the conventional optimization methods, the conditions are optimized by changing only one parameter and keeping the others unchanged ("one at a time" method). Apart from being time consuming, this method may not lead in accurate results. In most studies, only one or two parameters involved in biodegradation are examined. For instance, Dehghani et al. (26) optimized some parameters influencing alachlor biodegradation such as pH. They examined these factors individually, changing one factor and keeping the others unchanged (27).

Response Surface Methodology naturally includes a proper number of repetitions, yet the data analysis showed that the additional repetitions conducted in this study could lead to further accuracy. Therefore, it seems that the use of RSM and examination of more parameters is a better solution for the optimization of biodegradation conditions (29). Salinity of 14% determined at the optimization stage was in accordance with the salinity found for the optimum growth conditions of *Halomonas* sp. TBZ3 (7.5% - 20%). The optimum pH of 6.74 found for biodegradation of para-amino acetanilide was also in accordance with the optimum growth pH of the *Halomonas* sp. TBZ3 (pH 6 - 9) (16).

In conclusion, our findings showed that *Halomonas* sp. TBZ3 isolated from the Urmia Lake of Iran is able to degrade PAA. Optimum biodegradation conditions were determined through RSM as an efficient and accurate method. These optimum conditions were in accordance with optimum growth conditions and increased biodegradation rate remarkably. Further studies regarding potential of *Halomonas* sp. TBZ3 in biodegradation of other aromatic compounds are suggested.

## Authors' Contributions

Nader Hajizadeh and Yousof Sefidi Heris analyzed the samples and wrote the original manuscript. Sepideh

Zununi Vahed isolated TBZ3 from the Urmia Lake and performed the characterization experiments. DNA-DNA hybridization experiment was conducted by DSMZ (Germany) and BCCM/LMG (Belgium). G + C content were carried out by DSMZ. Javad Vallipour participated in optimization and performed the statistical analysis. Mohammad Amin Hejazi and Sayyed Mahdi Golabi participated in interpretation of the data and also editing of the manuscript. Mohammad Saeid Hejazi and Karim Asadpour-Zeynali designed the methods and supervised the project. Mohammad Saeid Hejazi edited the manuscript. All authors read and approved the final manuscript.

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## References

1. Khajamohiddin S, Repalle ER, Pinjari AB, Merrick M, Siddavattam D. Biodegradation of aromatic compounds: an overview of meta-fission product hydrolases. *Crit Rev Microbiol*. 2008;**34**(1):13-31.
2. Zhang MM, Chen WM, Chen BY, Chang CT, Hsueh CC, Ding Y, et al. Comparative study on characteristics of azo dye decolorization by indigenous decolorizers. *Bioresour Technol*. 2010;**101**(8):2651-6.
3. Firmino PI, da Silva ME, Cervantes FJ, dos Santos AB. Colour removal of dyes from synthetic and real textile wastewaters in one- and two-stage anaerobic systems. *Bioresour Technol*. 2010;**101**(20):7773-9.
4. O'Neil MJ. *The Merck index: an encyclopedia of chemicals, drugs, and biologicals*. Cambridge, UK: Royal Society of Chemistry; 2013.
5. Elisangela F, Andrea Z, Fabio DG, de Menezes Cristiano R, Regina DL, Artur C. Biodegradation of textile azo dyes by a facultative *Staphylococcus arlettae* strain VN-11 using a sequential microaerophilic/aerobic process. *Int Biodeteriorat Biodegrad*. 2009;**63**(3):280-8.
6. Jafari N, Kasra-Kermanshahi R, Soudi MR, Mahvi AH, Gharavi S. Degradation of a textile reactive azo dye by a combined biological-photocatalytic process: *Candida tropicalis* Jks2 -Tio2/Uv. *Iranian J Environ Health Sci Eng*. 2012;**9**(1):33.
7. Abari AH, Emtiazi G, Ghasemi SM, Roghanian R. Isolation and Characterization of a Novel Toluene-Degrading Bacterium Exhibiting Potential Application in Bioremediation. *Jundishapur J Microbiol*. 2013;**6**(3):256-61.
8. Moghadam MS, Ebrahimipour G, Abtahi B, Ghasempour A. Isolation, Identification and Optimization of Phenanthrene Degrading Bacteria From the Coastal Sediments of Nayband Bay. *Jundishapur J Microbiol*. 2013;**6**(9)
9. Mohsenzadeh F, Chehregani Rad A, Akbari M. Evaluation of oil removal efficiency and enzymatic activity in some fungal strains for bioremediation of petroleum-polluted soils. *Iranian J Environ Health Sci Eng*. 2012;**9**(1):26.
10. Balamurugan B, Thirumarimurugan M, Kannadasan T. Anaerobic degradation of textile dye bath effluent using *Halomonas* sp. *Bioresour Technol*. 2011;**102**(10):6365-9.
11. Oie CS, Albaugh CE, Peyton BM. Benzoate and salicylate degradation by *Halomonas campisalis*, an alkaliphilic and moderately halophilic microorganism. *Water Res*. 2007;**41**(6):1235-42.
12. Arahal DR, Ventosa A. The family Halomonadaceae. In: Dworkin M, Falkow S, editors. *The prokaryotes*. 3 ed. Springer; 2006. pp. 811-35.
13. Margesin R, Schinner F. Potential of halotolerant and halophilic microorganisms for biotechnology. *Extremophiles*. 2001;**5**(2):73-83.
14. Ventosa A, Nieto JJ, Oren A. Biology of moderately halophilic aerobic bacteria. *Microbiol Mol Biol Rev*. 1998;**62**(2):504-44.

15. Garcia MT, Mellado E, Ostos JC, Ventosa A. Halomonas organivorans sp. nov., a moderate halophile able to degrade aromatic compounds. *Int J Syst Evol Microbiol.* 2004;**54**(Pt 5):1723-8.
16. Vahed SZ, Forouhandeh H, Hassanzadeh S, Klenk HP, Hejazi MA, Hejazi MS. Isolation and characterization of halophilic bacteria from Urmia Lake in Iran. *Mikrobiologiia.* 2011;**80**(6):826-33.
17. De Ley J, Cattoir H, Reynaerts A. The quantitative measurement of DNA hybridization from renaturation rates. *Eur J Biochem.* 1970;**12**(1):133-42.
18. Huss VA, Festl H, Schleifer KH. Studies on the spectrophotometric determination of DNA hybridization from renaturation rates. *Syst Appl Microbiol.* 1983;**4**(2):184-92.
19. Ezaki T, Hashimoto Y, Yabuuchi E. Fluorometric deoxyribonucleic acid-deoxyribonucleic acid hybridization in microdilution wells as an alternative to membrane filter hybridization in which radioisotopes are used to determine genetic relatedness among bacterial strains. *Int J Sy Bacteriol.* 1989;**39**(3):224-9.
20. Larcher S, Yargeau V. Biodegradation of sulfamethoxazole by individual and mixed bacteria. *Appl Microbiol Biotechnol.* 2011;**91**(1):211-8.
21. Lin C, Gan L, Chen ZL. Biodegradation of naphthalene by strain *Bacillus fusiformis* (BFN). *J Hazard Mater.* 2010;**182**(1-3):771-7.
22. Arulazhagan P, Vasudevan N. Role of a moderately halophilic bacterial consortium in the biodegradation of polyaromatic hydrocarbons. *Mar Pollut Bull.* 2009;**58**(2):256-62.
23. Ghodake G, Jadhav S, Dawkar V, Govindwar S. Biodegradation of diazo dye Direct brown MR by *Acinetobacter calcoaceticus* NCIM 2890. *Int Biodeteriorat Biodegrad.* 2009;**63**(4):433-9.
24. Olukanni OD, Osuntoki AA, Kalyani DC, Gbenle GO, Govindwar SP. Decolorization and biodegradation of Reactive Blue 13 by *Proteus mirabilis* LAG. *J Hazard Mater.* 2010;**184**(1-3):290-8.
25. Srinivasan SV, Murthy DV. Statistical optimization for decolorization of textile dyes using *Trametes versicolor*. *J Hazard Mater.* 2009;**165**(1-3):909-14.
26. Dehghani M, Nasserli S, Zamanian Z. Biodegradation of alachlor in liquid and soil cultures under variable carbon and nitrogen sources by bacterial consortium isolated from corn field soil. *Iranian J Environ Health Sci Eng.* 2013;**10**(1):21.
27. Bryant J, Brereton DA, Gill CO. Implementation of a validated HACCP system for the control of microbiological contamination of pig carcasses at a small abattoir. *Can Vet J.* 2003;**44**(1):51-5.
28. Chen CY, Kuo JT, Cheng CY, Huang YT, Ho IH, Chung YC. Biological decolorization of dye solution containing malachite green by *Pandora pulmonicola* YC32 using a batch and continuous system. *J Hazard Mater.* 2009;**172**(2-3):1439-45.
29. Khataee AR, Zarei M, Ordikhani-Seyedlar R. Heterogeneous photocatalysis of a dye solution using supported TiO<sub>2</sub> nanoparticles combined with homogeneous photoelectrochemical process: Molecular degradation products. *J Mol Catalys Chem.* 2011;**338**(1):84-91.